

1 TEMPORAL VARIABILITY OF CHLORINATED VOLATILE
2 ORGANIC COMPOUND VAPOR CONCENTRATIONS IN A
3 RESIDENTIAL SEWER AND LAND DRAIN SYSTEM
4 OVERLYING A DILUTE GROUNDWATER PLUME

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10

11 **Abstract**

12 Some subsurface sewer and land drain networks will facilitate the migration of chlorinated
13 volatile organic compounds (CVOCs) from dissolved contaminant groundwater plumes to indoor
14 air. As this vapor intrusion (VI) pathway has only recently been documented, guidance for
15 evaluating it, including recommendations for timing, frequency, duration and location for vapor
16 sampling in subsurface piping networks is non-existent. To address this gap, a three-year
17 investigation of CVOC concentrations from land drains, storm drains, and sanitary sewers was
18 undertaken in a neighborhood overlying a large-scale dissolved chlorinated VOC (CVOC)
19 groundwater plume. Vapor sampling included the collection of grab (time-discrete) samples from
20 up to 277 manholes, hourly grab sampling from three manhole locations, and 24-h duration
21 collection during week-long sampling from 13 land drain and sewer manholes. The spatial
22 distribution of vapor and water concentrations and the temporal variations in the vapor values
23 observed in this study suggest that week-long vapor sampling conducted at different times of the
24 year and with samples collected at manhole locations overlying and outside a dissolved plume
25 might be needed to ensure robust VI pathway assessment at other sites. These findings are

26 expected to be of relevance to regulatory agencies involved in the development of current or
27 future VI pathway assessment guidance.

28

29 **Keywords:**

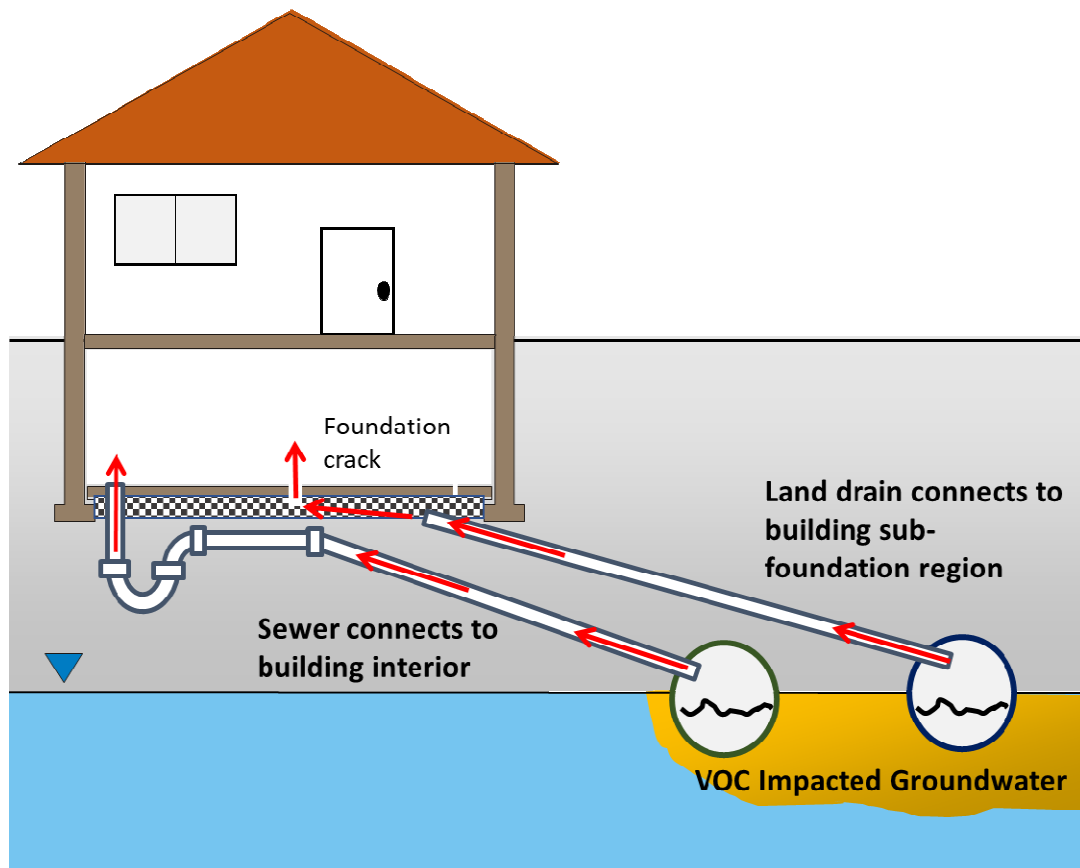
30 Vapor intrusion, indoor air, chlorinated volatile organic chemicals, groundwater, sewers, land
31 drains

32

33 **1.0 Introduction**

34

35 Vapor intrusion (VI) field studies have shown that indoor air in buildings connected to sewer and
36 land drain systems (sub-surface drainage systems that prevent water accumulation beneath
37 building foundations) can be impacted by volatile organic chemical (VOC) vapors present in the
38 sewers and land drains (Guo *et al.*, 2015; McHugh *et al.*, 2017; McHugh and Beckley, 2018;
39 Pennell *et al.*, 2013; Riis *et al.*, 2010; Roghani *et al.*, 2018). This often occurs when
40 contaminated groundwater enters the sewer or land drain system, as shown in Figure 1. In these
41 cases, VOC contaminants volatilize and migrate along the piping headspace and finally enter
42 buildings either via a direct connection to indoor air (sewer in Figure 1); or through the sub-
43 foundation region and foundation cracks (land drain system in Figure 1). When such VI
44 pathways exist, VI impacts can occur to buildings that are connected to the contaminated
45 groundwater entry point, but do not overlie dilute VOC groundwater plumes (Riis *et al.*, 2010).
46 As a result, VI risk assessments need to consider this “pipe-flow” VI pathway in addition to the
47 conventional “soil VI” pathway where chemical vapors migrate upward from groundwater
48 plumes through soil and then into a building (Guo *et al.*, 2015).



49

50 Figure 1. Conceptual illustration of sewer and land drain vapor intrusion pathways.

51

52 Although the evaluation of alternative and preferential VI pathways is mentioned in federal and
 53 state regulatory guidance (ITRC, 2007; NJDEP, 2013; USEPA, 2015), there is little guidance on

54 how to specifically identify or assess their VI risks. The lack of available guidance is, in part,

55 because these VI pathways have only recently been recognized and documented (Riis *et al.*,

56 2010; Pennell *et al.*; 2013; Guo *et al.*, 2015; McHugh *et al.*, 2017; McHugh and Beckley, 2018).

57 While approaches for assessing potential indoor air impacts from VOCs in sewers and drains

58 have yet to be developed or validated, guidance is likely to include requirements for source vapor

59 concentration characterization and extrapolation of inhalation exposure using empirical relations

60 or mathematical models. Thus, guidance for the characterization of VOC vapor concentrations

61 in sewers, land drains, and other subsurface piping will be needed, including specification of
62 sample collection and analysis methods and the time, duration, and frequency of sampling.

63

64 The presence of VOC vapors in subsurface piping networks has been reported in studies that
65 discuss odor management in sewer networks, and most of these studies have focused on specific
66 analytical constituents and their concentration levels (Corsi *et al.*, 1995; Quigley and Corsi,
67 1995; Corsi and Quigley, 1996; Yeh *et al.*, 2011; Huang *et al.*, 2012; Wang *et al.*, 2012a, 2012b,
68 2015). However, the temporal variability of VOC vapor concentrations in subsurface piping
69 networks is not well-understood. Only a limited number of studies have investigated this topic,
70 and their observations and conclusions were based on VOC vapor monitoring either from limited
71 sampling locations or for short time period. Quigley and Corsi (1995) found weekday/weekend
72 trends for three aromatic compounds in 17 sewer manholes during four 24-h sampling events,
73 Sivret. *et al.* (2017) observed up to 10x diurnal VOC vapor concentration changes in a pump
74 station wet well, and Roghani *et al.* (2018) reported over 100x changes in trichloroethylene
75 (TCE) concentrations in two sewer manholes adjacent to a groundwater plume over a two-year
76 period.

77

78 The observations from past studies are informative but not sufficient to create broadly applicable
79 guidance for characterizing VOC vapor concentrations in subsurface piping networks for use in
80 VI pathway risk assessment. Thus, this study was undertaken to address this gap through high-
81 and low-frequency sampling of chlorinated VOC (CVOC) vapors in land drains, storm drains,
82 and sanitary sewers located in a neighborhood overlying a large-scale dissolved CVOC
83 groundwater plume. Sampling was conducted over a period of about three years with the

84 sampling efforts changing as more was learned about the levels and dynamics of vapor
85 concentrations in the system. The sampling included multi-season synoptic collection of
86 instantaneous grab samples from up to 277 manholes, hourly grab samples from two land drain
87 locations and a sanitary sewer manhole, and multi-season week-long collection of 24-h duration
88 samples from 13 land drain manholes.

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90 **2.0 Methods**

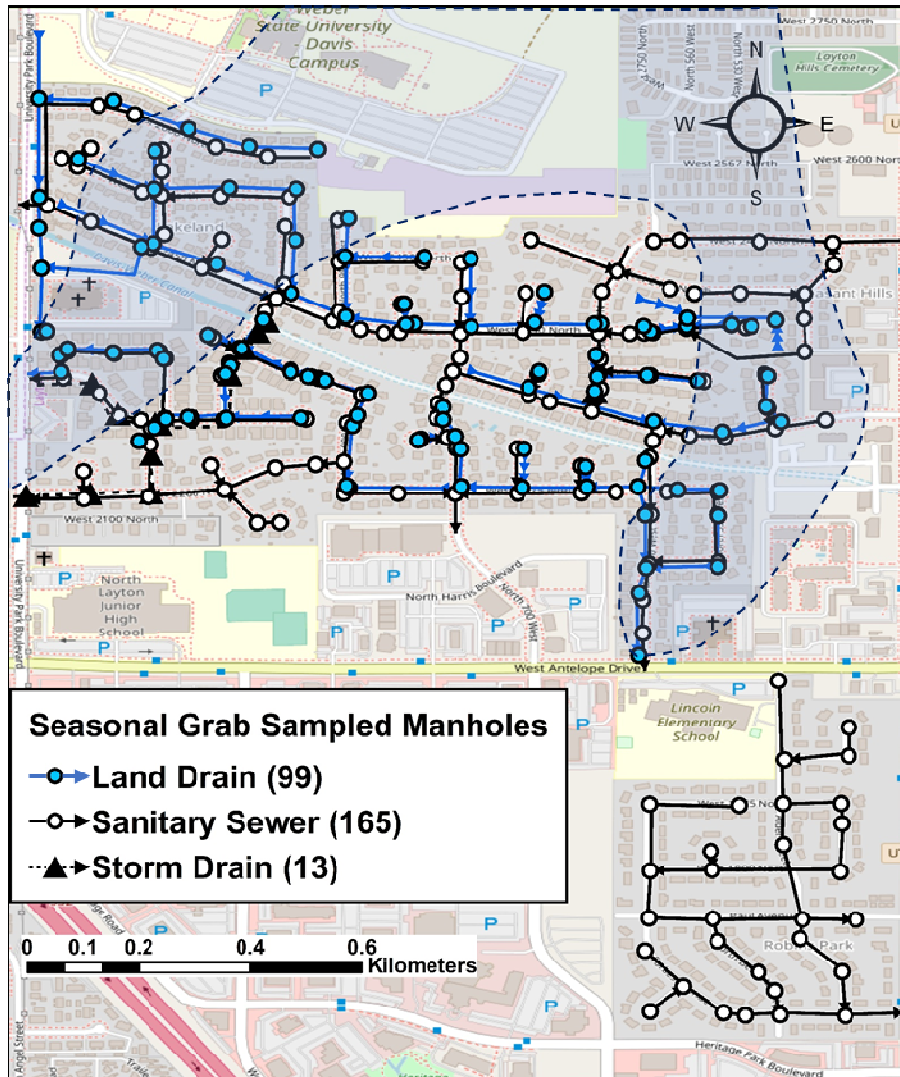
91 **2.1 Study Site**

92

93 Air and water sampling were conducted over an approximately 1 km² residential area adjacent to
94 Hill Air Force Base, UT. This area overlies a shallow dilute CVOC groundwater plume and
95 throughout the study area there are land drain and storm water and sanitary sewer networks. TCE
96 is the primary VI contaminant of concern within the study area where TCE dissolved
97 groundwater concentrations range from approximately 5 ug/L to 100 ug/L (Hill Air Force Base,
98 2005). The land drain system has been previously confirmed as the source of CVOC indoor air
99 impacts for one intensely studied residence (Guo *et al.*, 2015). The dissolved plume boundaries
100 and 277 sampled manhole locations are presented in Figure 2.

101

102



103
 104 Figure 2. Study area and locations of sampled manholes. The shaded area bounded by the dashed
 105 line delineates the dissolved TCE groundwater plume. Arrows indicate direction of water flow
 106 in the subsurface piping networks.

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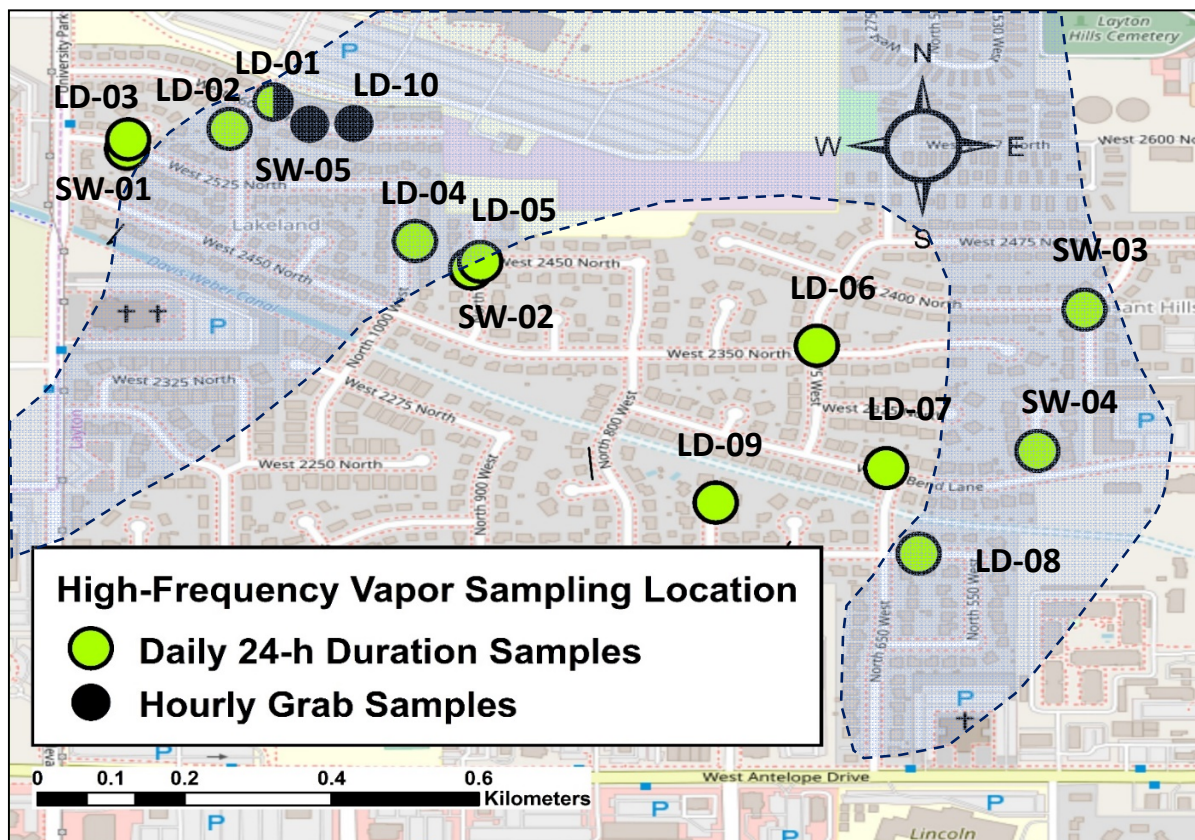
108 2.2 Sample Collection Summary

109

110 Samples collections were performed from January 2016 to January 2019, through the following
 111 activities:

- 112 1) Multi-season grab sampling (January 2016 to April 2017): five synoptic grab sampling
113 events were performed to characterize the spatial distribution of CVOC vapors in the
114 subsurface piping networks and to assess seasonal variability. Each event included vapor
115 sampling from up to 277 of the manholes shown in Figure 2 (165 sewer manholes, 99
116 land drain manholes, and 13 storm drain manholes). Since vapor phase VOCs in
117 subsurface piping networks are often the result of contaminated groundwater infiltration,
118 grab sampling of water from land drain manholes was also performed along with the
119 vapor sampling when water was present. These data are useful for assessing the value of
120 water sampling as another line of evidence for VOC characterization in subsurface piping
121 networks.
- 122 2) Hourly high-frequency grab sampling (September 2017 to March 2018): hourly sampling
123 was conducted over five months in the two land drain manholes and one sanitary sewer
124 manhole shown in Figure 3 to provide initial insight into shorter-term temporal variability
125 in CVOC vapor concentrations. All three were adjacent to the residence having a
126 confirmed pipe-flow VI alternative pathway from the land drain network.
- 127 3) Daily high-frequency sampling (March 2018 to January 2019): A total of six week-long
128 sampling events covering multiple seasons and involving the collection of daily 24-h
129 samples were performed using the 13 manholes (9 land drain, 5 sanitary sewer, and 1
130 sanitary sewer/storm drain combination) shown in Figure 3. These locations were
131 selected based on multi-season grab sampling results, with the intent of including
132 locations with a range of concentrations and temporal variabilities.

133



134
 135 Figure 3. Locations where hourly (black) and daily 24-h duration (green) vapor samples were
 136 collected for extended sampling periods. LD = land drain manhole; SW =sanitary sewer
 137 manhole. SW03 is a sanitary sewer/storm drain combination manhole.

138

139 2.3 Vapor Sample Collection and Analysis Methods

140

141 *Multi-season grab samples.* Manhole vapor samples were collected using a method similar to
 142 that described in McHugh *et al.* (2017). A vacuum box sampler was used to draw vapor samples
 143 (minimum 500 mL) into a Tedlar bag via weighted nylon tubing inserted through vent holes in
 144 the manhole covers. If vent holes were not present, the cover was opened just enough to allow
 145 passage of the sampling tubing. The distal end of the weighted tubing was inserted to a depth

146 approximately 0.3 m above the base of the manhole or manhole water level. The vapor samples
147 were analyzed on-site using an SRI gas chromatograph equipped with a dry electron capture
148 detector (GC/DELCD) (SRI instrument, CA) , and the minimum detection level (MDL) for TCE
149 analysis by this method was 1.5 ppb_v. The GC/DELCD was calibrated daily prior to sample
150 collection and calibration checks and duplicate vapor samples were analyzed every 10 sample
151 injections for QA/QC purposes. The average relative percentage differences between duplicate
152 samples was 26.9%.

153

154 *Hourly high-frequency grab sampling.* Hourly vapor grab samples were collected directly onto
155 the GC using an external pump, autosampler, and permanent nylon and stainless-steel sampling
156 lines extending to each manhole. Permanent sampling lines were installed to a depth 0.3 m
157 above the manhole base or water level. Samples were analyzed real-time using an SRI GC
158 equipped with an electron capture detector (ECD). The minimum detection limit for TCE was
159 1.5 ppb_v. The GC/ECD was calibrated approximately every 4 weeks during the sample collection
160 period.

161

162 *Daily 24-h duration samples.* 24-h duration samples were collected daily on multi-bed sorbent
163 tubes comprised of Tenax-GR and Carboxen-569 sorbents. The vapor samples were collected
164 using a customized sampler which was suspended in the manhole approximately 0.3 to 0.5-m
165 above the base of the manhole or water level. The sampler pulled vapor through each sorbent
166 tube at a controlled flowrate (about 50 mL/min) using a Gilian LFS-113 air pump (Sensidyne,
167 FL). The flowrate for each pump was calibrated before and after each 24-h tube sample
168 collection using a Sensidyne Gilibrator-2 bubble flowmeter (Sensidyne, FL). Flowrate variation

169 over a 24-h period was typically less than 5% and never exceeded 10%. Sorbent tubes were
170 analyzed using a Markes Ultra auto-sampler and Markes Unity thermal desorber (Markes
171 International, UK) connected to an HP5890 gas chromatograph equipped with a Restek 60 m
172 Rxi-5 capillary column and an HP5972 mass spectrometer. Samples were analyzed using
173 selective ion mode (SIM). The 24-h average CVOC concentration was calculated based on the
174 CVOC mass loading for sorbent tube and the vapor sample volume. The minimum TCE
175 detection level was 0.07 ppb_v. Duplicate samples were collected in manhole LD-02 and SW-03
176 and the variations in concentrations for duplicate samples and duplicate analyses were less than
177 30%.

178

179 **2.4 Water Sample Collection and Analysis**

180

181 Water samples were collected from land drain manholes and selected storm drain manholes
182 where possible during the area-wide seasonal grab sampling events. Samples were collected
183 from each manhole in 40 mL volatile organic analysis (VOA) vials, which contained 0.5 mL 2%
184 hydrochloric acid for preservation. All samples were stored at 4 °C and shipped to Arizona State
185 University for headspace analysis within two weeks of sample collection. An SRI GC/DELCD
186 was used for sample analysis with the minimum detection level of 0.7 µg/L for TCE. Calibration
187 checks and duplicate vapor samples were analyzed every 10 sample injections for QA/QC
188 purposes. The average relative percentage differences between duplicate samples was 21.6 %.

189

190 **3.0 Results and Discussion**

191

192 **3.1 TCE vapor and water concentrations spatial distributions**

193 Five area-wide synoptic sample collection events were conducted from early 2016 to mid-2017.

194 The first event (January 2016) included 82 manhole locations. As knowledge of the manhole
195 system and the ability to differentiate types of manholes improved, all accessible manholes
196 within the area were being sampled by August 2016.

197

198 TCE vapors were detected throughout the land drain, storm drain, and sanitary sewer network.

199 The results of all synoptic sampling events can be found in Supplemental Information Figure S1-

200 S4. Figure 4 provides an overview of the range of TCE vapor concentrations detected and how

201 that changed over the five multi-season synoptic sampling events. In this figure, TCE vapor

202 concentration distributions are presented in four concentration categories which ranges from less

203 than 4 ppb_v to over 400 ppb_v. To provide some context for these concentrations, published indoor

204 air screening levels for TCE range from about 0.09 – 0.4 ppb_v (e.g., MADPH 2017, USEPA

205 2019), with the lower level based on a 10⁻⁶ risk level and the upper based on 10⁻⁵ risk level, with

206 both also considering non-cancer risks. Manhole vapor concentrations were found to be 100x

207 and 10x greater than the indoor air screening level of 0.4 ppb_v (USEPA, 2019) in

208 approximately 10 % and 40% of manhole sampling locations, respectively. For context, indoor

209 air TCE concentrations in a study house located in this area were about 1% - 2% of the nearby

210 land drain vapor concentrations when the house was under-pressurized (Guo *et al.*, 2015; Holton

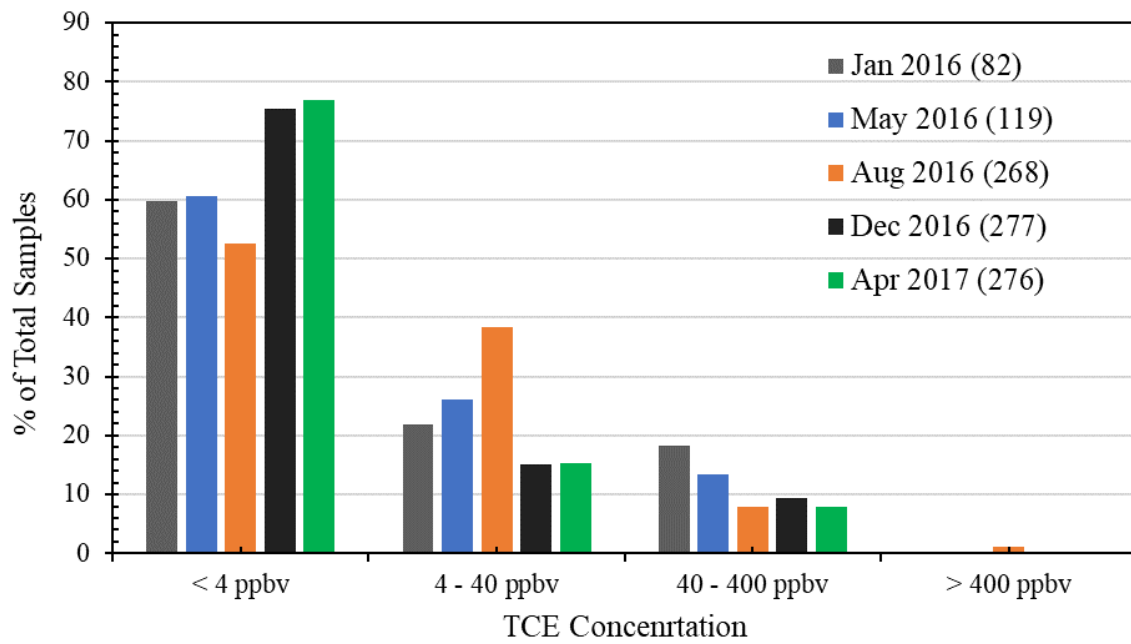
211 *et al.*, 2015) . Thus, residences near the higher-level manhole TCE vapor concentrations

212 measured in this study could be at risk of VI impact above the 0.4 ppb_v indoor air screening

213 level, but only if there are piping conduits connecting their homes to the land drain system.

214

215 One important observation from these synoptic sampling results is that the presence or
216 concentrations of TCE in the piping networks cannot be anticipated by groundwater plume data.
217 The poor correlation can be seen in Figures 5 and 6, which present the maximum TCE vapor and
218 water sample concentrations from the five synoptic sampling events superimposed on a map
219 showing the extent of the groundwater plume. About half of the locations where vapor
220 concentrations were >40 ppb_v were located outside of the groundwater plume boundary,
221 indicating that the piping networks were a conduit for dissolved and vapor-phase CVOC
222 transport to areas outside the groundwater plume. Although it was difficult to identify the exact
223 locations where groundwater entered the subsurface piping networks, TCE liquid samples were
224 all above 0.7 µg/L in the high-TCE-vapor-concentration-level manholes that were located
225 outside TCE groundwater plume boundary. This suggests that the migration of infiltrated
226 groundwater along the subsurface conduit's flow pathway is the primary mechanism for VOC
227 migration outside of the groundwater plume boundary. Thus, it is important that any future VI
228 pathway assessment guidance recommend sampling in subsurface piping networks beyond the
229 boundaries of dissolved groundwater plumes, particularly, when the depth of subsurface piping
230 networks is close to or deeper than groundwater table.
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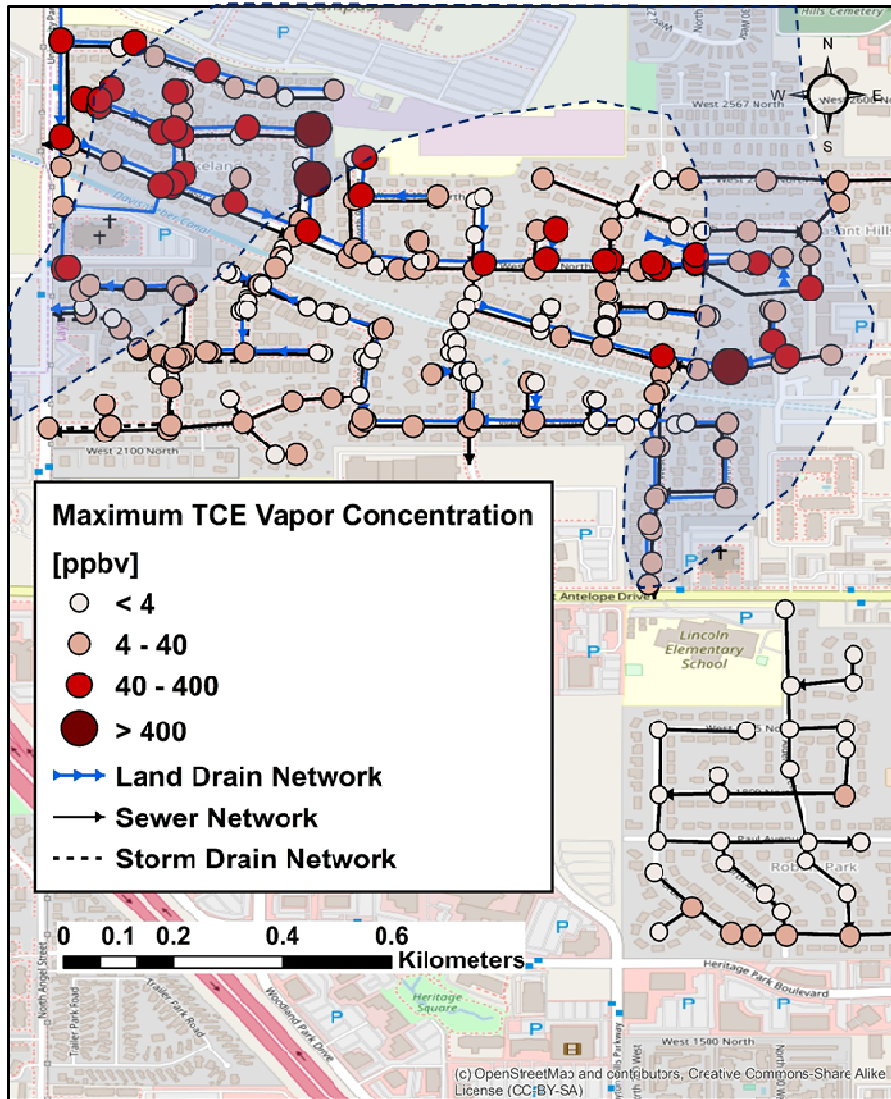
233 Figure 4. TCE manhole vapor concentration summary of five seasonal synoptic sampling events,
 234 categorized relative to a 0.4 ppbv indoor air screening level. Numbers of sampled manholes for
 235 each event are shown in brackets.

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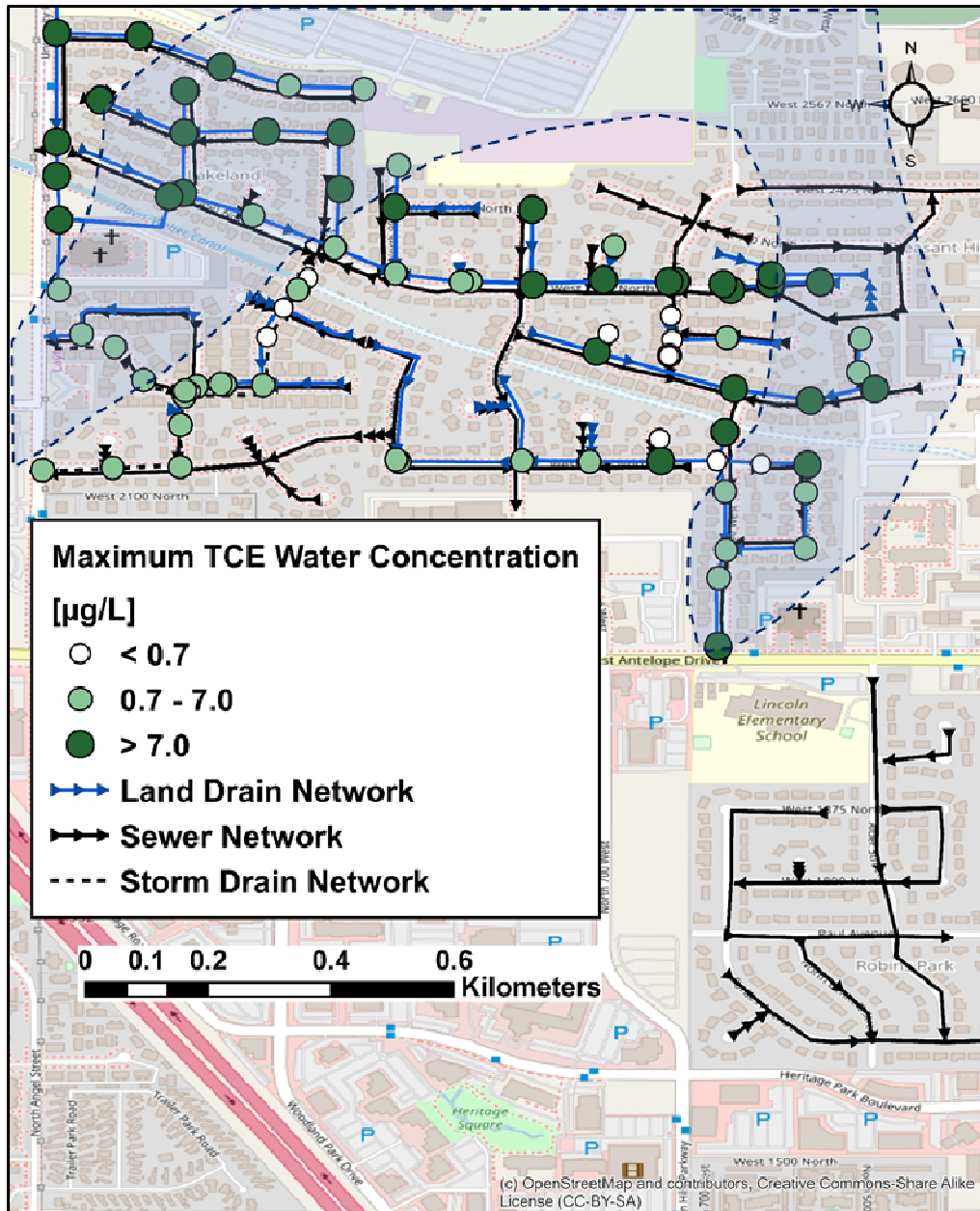


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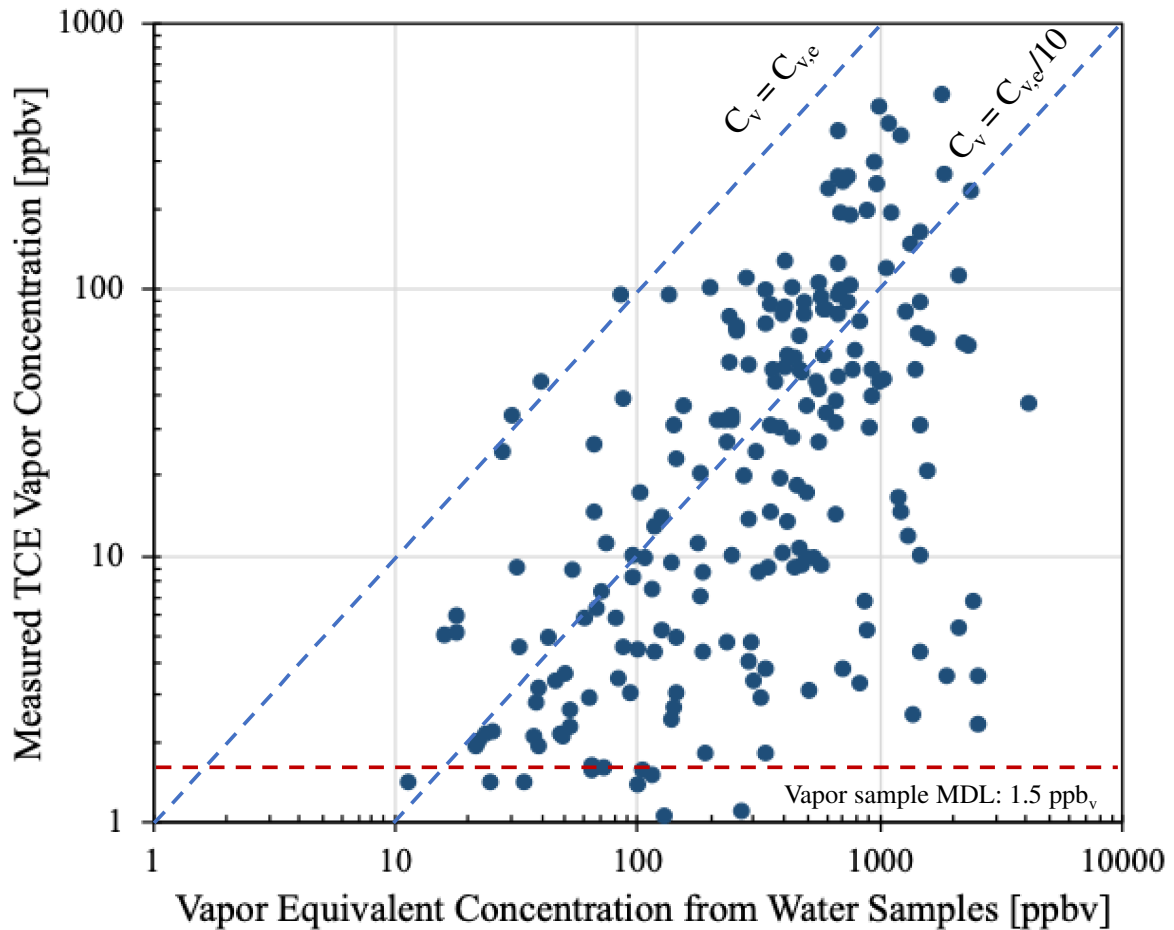
242 Figure 5. Maximum TCE concentrations in vapor samples collected from manhole headspace
 243 sampled during the five quarterly synoptic surveys, categorized relative to a 0.4 ppbv indoor air
 244 screening level. The shaded area indicates the extent of the TCE groundwater plume.
 245

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Figure 6. Maximum TCE concentrations in water samples collected from land drain manholes during the five quarterly synoptic surveys. The shaded area indicates the extent of the TCE groundwater plume.



255

256 Figure 7. Vapor equivalent concentration ($C_{v,e}$) vs. measured vapor concentration (C_v) for water
 257 and vapor samples collected in the same manhole. The Dimensionless Henry's Law Constant
 258 used in these calculations was 0.4 L-H₂O/L-vapor (USEPA, 2019).

259

260 VI guidance documents, from federal to state, all recognize dissolved VOC concentration in

261 groundwater as one important line of evidences for VI risk assessment, since dissolved water

262 concentrations can be used to predict vapor and indoor air concentrations, using the assumption

263 of local equilibrium. Thus, we examined the correlation between TCE concentrations in water

264 and vapor samples collected from the same manholes to evaluate the value of water sample

265 collection in VI pathway investigation. The results are presented in Figure 7 where the measured

266 headspace TCE vapor concentration (C_v) is plotted vs. the vapor equivalent concentration ($C_{v,e}$)
267 for the water samples, calculated by multiplying the measured dissolved TCE concentration in a
268 water sample by the dimensionless Henry's Law Constant for TCE (0.4 L-H₂O/L-vapor;
269 USEPA, 2019). A total of 256 paired water and vapor samples are plotted in Figure 7. As can be
270 seen, the measured TCE vapor concentrations were less than 10% of $C_{v,e}$ for 70% of the samples,
271 suggesting that use of VOC concentrations from water samples will lead to over-prediction of
272 VOC vapor concentrations when a simple local equilibrium assumption is applied. Corsi and
273 Quigley (1996) identified headspace ventilation rate, water flowrates and the water flow
274 conditions in manholes (fully submerged, partially submerged pipeline or water drops) as critical
275 factors that affect VOC migration rate from liquid to vapor phase in piping networks. Therefore,
276 these factors should be evaluated if VOC liquid sample concentrations were used for VI risk
277 characterization. However, sewer ventilation rates and water flow rates in pipelines could not be
278 easily quantified, and accurate measures of these often require intensive efforts, such as tracer
279 releasing. As such, it is best to collect and analyze vapor samples from subsurface piping
280 networks, rather than water samples, for VI pathway assessment.

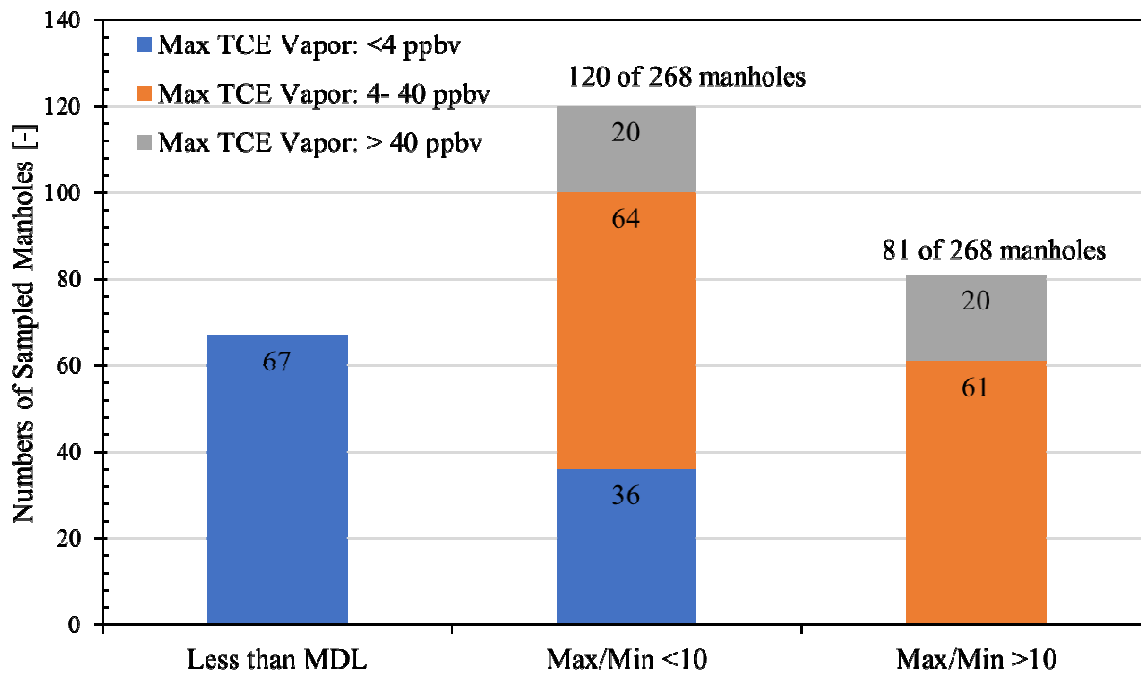
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282 **3.2 Temporal Variability in Multi-Season Grab Sample Concentrations**

283

284 The temporal changes in the multi-season grab sample results were assessed by looking at the
285 maximum/minimum concentration ratio at each of the 268 locations where at least three
286 sampling events occurred. Any sample result that was non-detect was assigned a value of one-
287 half the MDL (0.75 ppb_v) in these calculations. The results were then parsed into the three
288 groups shown in Figure 8 and discussed below:

- 289
- Group I: Locations where TCE manhole headspace concentrations were consistently
290 below the MDL (67 of 268 manholes). These are locations where the temporal variability
291 could not be assessed with the data and the concentrations at these locations are unlikely
292 to cause VI indoor air impacts above a 0.4 ppb_v TCE indoor air screening level.
 - Group II: Locations where TCE vapor concentrations were measured above the MDL at
293 least once, at relatively stable levels as their maximum/minimum TCE vapor
294 concentration ratios were <10x. This group includes 120 of 268 manholes, and of those,
295 there were 64 locations where the maximum concentration was between 10x and 100x of
296 a 0.4 ppb_v indoor air screening level.
 - Group III: Locations where significant changes in concentration occurred as the
298 maximum/minimum TCE vapor concentration ratios were >10x. This set includes about
299 30% (81 of 268) of the sampled manholes. Most of these locations (61) had contrasting
300 concentrations that might be judged to be both of concern (>10x a 0.4 ppb_v screening
301 level) and not of concern (<10x a 0.4 ppb_v screening level). The largest
302 maximum/minimum TCE vapor concentration ratio was >500x.
303
- 304



305

306 Figure 8. Summary of temporal TCE vapor concentration changes in multi-season grab sample
 307 results.

308

309 Overall, relatively stable vapor concentrations were observed at some locations and highly
 310 variable results were observed at others, without any way to anticipate the temporal variabilities
 311 or maximum concentration at any specific location. Of the Group III locations – those with the
 312 greatest changes between samples – the maximum concentration was measured during a winter
 313 sampling event at 21% of these manholes and the maximum concentration was measured in a
 314 summer sampling event at 72% of the manholes. This suggests that it would be prudent for
 315 future guidance to recommend multi-season sampling events when assessing potential VI
 316 impacts from subsurface piping networks.

317

318 **3.3 Real-time Hourly Sampling Results**

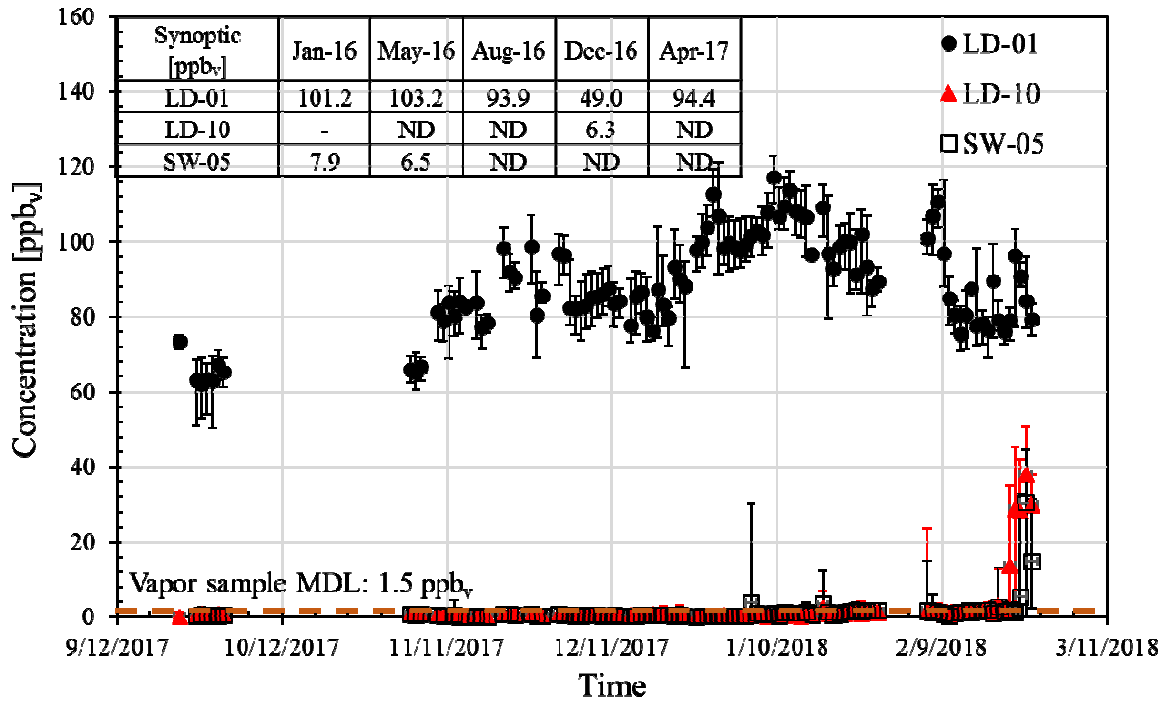
319 To assess if the changes observed in multi-season sampling results reflected long-term seasonal
320 changes or shorter-term (hourly to daily) vapor concentration fluctuations, hourly grab sampling
321 was conducted at selected manholes that had both consistent and highly variable multi-season
322 results. Hourly samples collected from LD-01, LD-10 and SW-05 (Figure 3) for about five
323 months (September 2017 to March 2018) were averaged for each day and plotted as presented in
324 Figure 9, showing also the maximum and minimum result from each 24-h period.

325 Manhole headspace TCE concentrations were consistently below the MDL for over 90% of the
326 sampling period in both LD-10 and SW-05, followed by spikes to 51 ppb_v and 45 ppb_v,
327 respectively, in early spring. This pattern is consistent with their multi-season sampling results:
328 at LD-10 and SW-05 the TCE headspace concentrations were <MDL for three of four events and
329 three of five events, respectively. In contrast the LD-01 concentrations were mostly in the 50 –
330 120 ppb_v range, with differences between daily maximum and minimum TCE vapor
331 concentration being <35% of the 24-h averaged TCE concentration values each day. LD-01
332 hourly TCE concentrations ranged from 50.3 ppb_v to 122.7 ppb_v with an averaged value of 89.9
333 ± 13.4 ppb_v (average ± standard deviation), which was consistent with the multi-season results
334 that ranged from 49 - 103 ppb_v from seasonal synoptic survey samples.

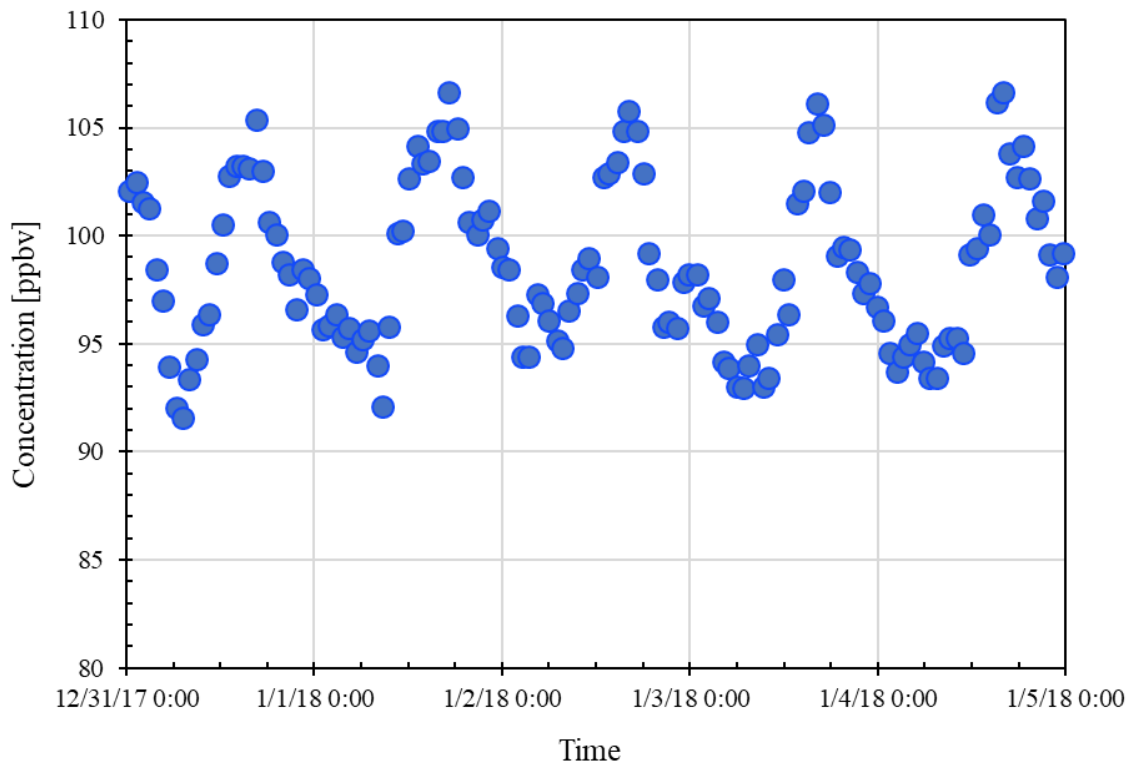
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336 To provide additional insight to short-term concentration variations, Figure 10 presents hourly
337 sample results vs. time for a five-day period at the LD-01 location. A diurnal pattern is evident
338 in the data with the TCE vapor concentrations reaching their highest level in late afternoon and
339 decreasing during the night. This short-term (24 h) variability in TCE vapor concentration was
340 not significantly different from the long-term (multi-season) variation. The ratio of daily

341 maximum/minimum concentrations was typically <1.2, while it was about 2 for the multi-season
 342 sampling data at LD-01.
 343



344
 345 Figure 9. 24-h averaged manhole headspace TCE concentrations at LD-01, LD-10 and SW-05
 346 (see Figure 2). Error bars denote the daily maximum and minimum values.
 347



348
 349 Figure 10. Diurnal behavior of TCE vapor concentrations in the LD01 manhole headspace.
 350

351 **3.4 24-hour Thermal Desorption Sampling Results**

352 To further assess the temporal variability in manhole headspace vapor concentrations, six week-
 353 long sampling events were conducted from March 2018 to January 2019. During each, 24-h
 354 time-integrated samples were collected from 13 manholes. The 13 manholes were selected based
 355 on their multi-season grab sampling results, with the goal of including locations with different
 356 patterns of results: two manhole locations where concentrations were consistently below the
 357 MDL (Group I in Figure 8); five manhole locations where concentrations varied by <10x (Group
 358 II in Figure 8); and six manhole locations where concentrations varied by more than 10x (Group
 359 III in Figure 8).

360

361 Table 2. Statistical summary of the week-long period 24-h sampling results with corresponding
 362 seasonal grab sampling results at each location.
 363

| Seasonal Variation | Manhole ID | TCE Vapor Concentration [ppb _v] | | | | | | | | |
|--------------------------------------|------------|---|---------|---------|---------|---------|--|---------|---|---------------------------------|
| | | Multi-Season Grab Sample Results | | | | | Weekly Averages of the 24-h Sample Results | | Averages Across the Six Week-Long Sampling Events | |
| | | Jan-16 | May-16 | Aug-16 | Dec-16 | Apr-17 | Maximum | Minimum | Max 24-h Value/Weekly AVG Value | Min 24-h Value/Weekly AVG Value |
| Group I: All < MDL | LD-08 | NA | NA | <MDL(s) | <MDL(s) | <MDL(s) | 0.1 | <MDL(w) | 3.2 | 0.27 |
| | LD-09 | NA | NA | <MDL(s) | <MDL(s) | <MDL(s) | <MDL(w) | <MDL(w) | 2.6 | 0.17 |
| Group II: <10x Multi-season Max/Min | LD-05 | 49.0 | 37.3 | 13.6 | 31.9 | 19.5 | 37.9 | 11.2 | 1.3 | 0.71 |
| | LD-01 | 101.2 | 103.2 | 93.9 | 49.0 | 94.4 | 65.6 | 29.9 | 1.4 | 0.65 |
| | LD-07 | NA | 191.0 | 103.5 | 79.8 | 88.9 | 94.4 | 42.8 | 1.4 | 0.60 |
| | SW-02 | NA | 3.0 | 2.1 | 5.0 | <MDL(s) | 0.6 | <MDL(w) | 3.0 | 0.29 |
| | LD-06 | NA | NA | 31.2 | 98.2 | 83.2 | 59.8 | 1.1 | 2.4 | 0.48 |
| Group III: >10x Multi-season Max/Min | SW-01 | NA | 23.9 | 136.7 | <MDL(s) | 36.7 | 78.4 | 0.4 | 2.3 | 0.54 |
| | LD-04 | NA | 2.5 | 410.0 | 39.0 | 14.4 | 7.9 | 0.1 | 2.6 | 0.31 |
| | SW-03 | <MDL(s) | <MDL(s) | 11.8 | <MDL(s) | <MDL(s) | 0.1 | <MDL(w) | 2.7 | 0.022 |
| | SW-04 | NA | NA | 9.1 | 2.9 | <MDL(s) | 0.9 | 0.1 | 2.9 | 0.19 |
| | LD-02 | NA | <MDL(s) | 1.9 | 385.7 | 55.3 | 198.8 | 1.9 | 2.4 | 0.24 |
| | LD-03 | 37.0 | 62.3 | 4.3 | 49.7 | 45.5 | 127.5 | 4.5 | 1.6 | 0.45 |

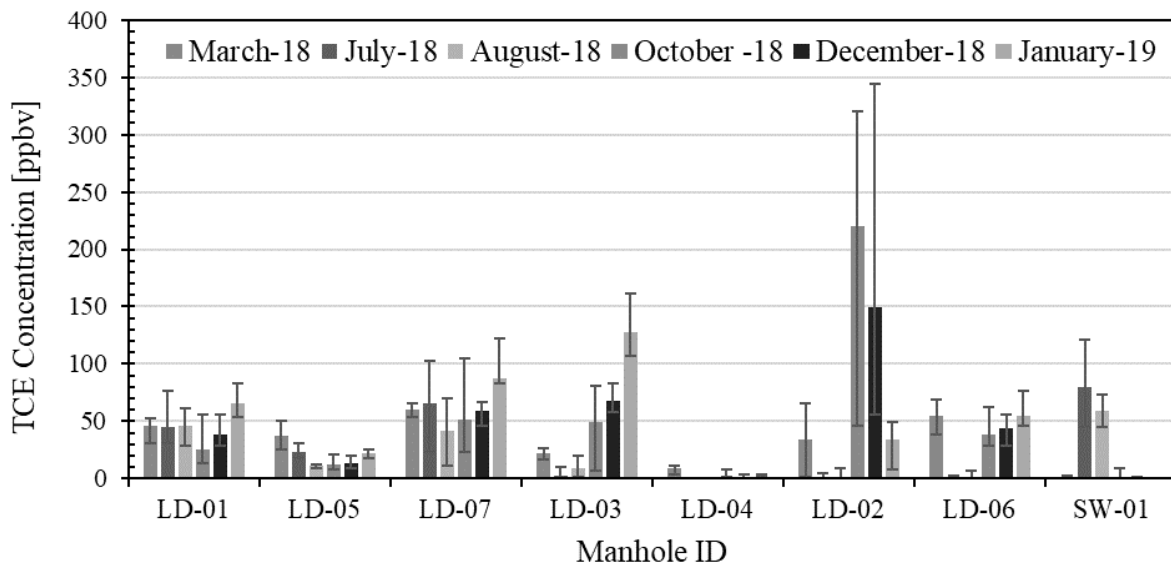
364
 365 NA – No sample available;
 366 MDL(s) – TCE detection limit for the synoptic samples: 1.5 ppb_v.
 367 MDL(w) – TCE detection limit for 24h samples: 0.07 ppb_v
 368

369 The results of this study are presented in Table 2 and Figure 11. A summary of the week-long
 370 period daily-sample results along with their multi-season grab sampling results are provided in
 371 Table 2. Figure 11 presents the averaged week-long sampling results for locations with
 372 concentrations >MDL, with the error bars spanning the maximum and minimum 24-h TCE vapor
 373 concentrations that were measured during each week-long sampling period.

374
 375 Collectively the results are mostly consistent with the synoptic and extended hourly sampling
 376 results. At some locations, the concentrations appear relatively temporally stable and were
 377 similar to grab sample, 24-h sample, and weekly-average results for those locations (e.g. LD-01,

378 -05, and -07). At those locations, grab samples collected at any time of the year would likely
 379 provide good insight to the concentrations, although increasing to weekly-average samples could
 380 decrease variability in sample results relative to grab or 24-h samples. At other locations (e.g.,
 381 LD-02 and -03), the 24-h and weekly-average results span a wide range, but encompassing
 382 values similar to the multi-season grab samples. At those locations, multi-season sampling
 383 would be needed to characterize the range of vapor concentrations at those locations, and grab,
 384 24-h, and weekly average samples would likely yield similar results. Then there are other
 385 locations (e.g., LD-06) where the multi-season grab samples suggested much less temporal
 386 variability than was revealed in the 24-h and weekly-average results or the maximum
 387 concentration detected in grab sampling was much greater than either 24-h sample or weekly-
 388 average results (e.g. 30x at LD-04).

389



390

391 Figure 11. The weekly averaged TCE headspace concentrations of 24-h samples with error bars
 392 spanning the maximum and minimum 24-h concentrations of each week-long sampling period.
 393

394

395 **4.0 Implication for VI Alternative Pathway Sampling in Sewers and**
396 **Other Subsurface Utility Conduits**
397

398 Overall, the following observations are supported by the data collected in this study:

- 399 • Diurnal concentration changes in hourly TCE vapor samples were less than 50% at one
400 intensely sampled location in this study. If concentration variations of this magnitude
401 about an average are of concern, the uncertainty in concentration results can be
402 minimized by collecting 24-h time-integrated samples.
- 403 • In our data set, individual 24-h average results ranged from 50% to 150% of the
404 calculated weekly-average at some locations (e.g., LD-01 and -07), but also varied to a
405 greater degree at other locations (e.g. LD-02 and -04). Thus, serious consideration
406 should be given to week-long sample durations rather than grab samples or 24-h sample
407 durations in designing alternate VI pathway assessment plans.
- 408 • Whether collecting grab, 24-h, or week-long samples, seasonal variability should be
409 expected. This was greater than daily or weekly variability at many locations at our study
410 site, so it is possible to measure concentrations of significance at some periods of the year
411 while seeing insignificant concentrations at others. For example, over 10x seasonal
412 variability was observed at 81 of 268 manholes in this study.
- 413 • Thus, multi-season synoptic events should be considered, as these are likely to provide
414 more confidence in characterizing vapor distributions in subsurface utilities than one-time
415 grab sampling events.
- 416 • Sampling location selection should not be overly constrained by dissolved plume
417 delineation as concentrations of significance have been observed in this and other studies
418 at locations outside of the dissolved plume footprint.

419 In brief, the results of this study suggest that robust alternate VI pathway sampling protocols
420 would typically include week-long samples collected at different times of the year with samples
421 collected at manhole locations overlying and outside the dissolved plume. Locations exterior to
422 the plume might be chosen based on connectivity and how flow occurs in the sewer and drainage
423 network, if that is known. It may be that week-long active vapor sampling at large numbers of
424 locations might be impracticable at sites with large dissolved plumes like our study site, so we
425 recommend that the utility and accuracy of passive sampling tools in sewer environments as
426 alternatives to active sampling be evaluated in future studies.

427

428 **Acknowledgements**

429 This research was funded by the U.S. Department of Defense, through Environmental Security
430 Technology Certification Program (ESTCP) Project ER-201501. Dr. Thomas McHugh and Lila
431 Beckley from GSI kindly provided their insights and useful suggestions throughout this research.

432

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